

Oxidative effects of glyphosate on the lipophobic intracellular environment in the microalgae *Chlorella vulgaris*

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Abstract: The studied hypothesis is that the herbicide glyphosate (GLY) can affect the oxidative balance in the hydrophobic intracellular medium in non-target *Chlorella vulgaris* cells. Analytical GLY and RoundUp (RUP) supplementation, affected the growth profile. A significant 42% decrease in the cellular biomass in stationary (St) phase was observed in cultures supplemented with either 5 μM of GLY or RUP, as compared to control cultures. The treatment with 0.3 μM of GLY generated non-significant effects on the oxidation rate of 2', 7' dichlorofluorescein diacetate (DCFH-DA), neither in exponential (Exp) nor in St phase of development, as compared to control cultures. However, the treatment with either 5 μM GLY or 0.3 and 5 μM RUP lead to a significant decrease in the DCFH-DA oxidation rate, as compared to control cultures. The lipid radical (LR^\bullet) generation rate, detected by Paramagnetic Resonance Spectroscopy (EPR), was significantly increased in the presence of RUP, in Lag and Exp phase of growth. The non-enzymatic antioxidants, α -Tocopherol (α -T) and β -Carotene (β -C), are aimed to protect membranes against the damage produced by the radical reactions. The content of β -C was not significantly affected, as compared to control cultures, by any of the treatments, in both growth phases of cellular development. The content of α -T was significantly decreased by the supplementation with either 0.3 or 5 μM of RUP or 5 μM GLY. The LR^\bullet/α -T ratio, used as indicator of the oxidative balance in the hydrophobic cellular media, was significantly different between samples obtained from control and RUP-exposed microalgae in both, Exp and St phase of development, with either 0.3 or 5 μM RUP. The data presented here showed evidence that suggested that oxidative balance in the hydrophobic environment was affected by either GLY or RUP.

Introduction

Glyphosate ($\text{C}_6\text{H}_{16}\text{NO}_5\text{PS}$) (GLY) is a broad spectra herbicide that is mainly used for weed control in crops of agricultural interest. Studies over the past decade in a variety of biological models, have shown that GLY and its commercial formulations are responsible for associated toxicity of variable magnitude. GLY enters aquatic ecosystems by accidental spraying, drifts or surface runoff (Altamar Ríos, 2007). GLY generates the death of the target systems by affecting their ability to synthesize proteins essential for survival. The activity of the enzyme 5-enolpiruvilshikimato-3-phosphate synthase, which is part of the metabolic pathway for production of aromatic amino acids, is inhibited by the GLY that behaves as an analogue of the second substrate

(phosphoenolpyruvate). Moreover, there is growing evidence that indicate that GLY (and its formulations) produce oxidative stress in cyanobacteria (Chen *et al.*, 2012), microalgae (Lipok *et al.*, 2010) and higher plants not natural targets for this herbicide (Ahsan *et al.*, 2008). Also, it has been suggested that the pathological effects in humans are due to a mechanism of free radical production due to the nonspecific chronic inflammation that it generates (Altamar Ríos, 2007). In the air, GLY is eliminated through reactions with hydroxyl radical ($\bullet\text{OH}$), and in the soil, it is degraded by the action of microorganisms, producing methylphosphonic ammonium acid (AMPA) and carbon dioxide (CO_2) (Altamar Ríos, 2007). In these reactions, free radicals and recombinant or fibrogenic cytokines are eliminated (Altamar Ríos, 2007). However, the real dimension of the magnitude of the oxidative effect generated by exposure to the herbicide is still a matter of deliberation.

The exposure of cultures of the algae *Chlorella kessleri*, to GLY at concentrations of 50 mg l^{-1} , much higher than the

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concentrations usually found in the environment, triggers oxidative stress, evidenced by the inhibition of its growth and morphological damage with disorganization of chloroplasts, lipid damage, increased protein concentration, and increased antioxidant defenses (GSH; superoxide dismutase, SOD and catalase, CAT) (Romero *et al.*, 2011). Choi *et al.* (2012) found that exposure to GLY affected chlorophyll-a fluorescence in five microalgal species, with certain interspecific differences in GLY sensitivity. Gomes and Juneau (2016), and Gomes *et al.* (2016a, 2016b) showed a strict relationship between GLY-induced oxidative stress and its deleterious effects on the photosynthetic processes. Also, Eker *et al.* (2006) and Cakmak *et al.* (2009) have reported that GLY can bind to different essential metals, causing a decrease in its bioavailability. Duke *et al.* (2012) have pointed out some contradictions on the mechanism of GLY to bind. The microalga *Chlorella vulgaris*, is frequently used for analysis of toxicity since it is easily grown under laboratory conditions. This alga is found in many freshwater aquatic environments, and it is seen as a good bioindicator of the presence of environmental contaminants. Qian *et al.* (2009) and Reno *et al.* (2016) have reported several studies on effects of the presence of contaminants, such as GLY and its commercial products, on *C. vulgaris*. More recently, Oстера *et al.* (2020) reported experimental evidence that strongly suggested that oxidative balance in the hydrophilic cellular environment is affected by GLY independently of the preparation used to administrate the product, even at low-medium concentrations currently in used. Moreover, it seems that the effect is reversible, either due to the magnitude of the herbicide-dependent damage or by the antioxidant activity endogenously activated. However, the reactive oxygen radical (ROS) and nitrogen radical species lead to damage to macromolecules, not only in the hydrophilic but in the hydrophobic medium, as well. The lipid peroxidation is understood as the oxidative deterioration of the polyunsaturated lipids and conduct to effects such as alterations in the structure, fluidity and permeability of cellular membranes. The term lipid radicals (LR^{\bullet}), generated by peroxidation, include alkyl, peroxy (ROO^{\bullet}) and alkoxy radicals. Non-enzymatic lipid soluble antioxidants, such as α -Tocopherol (α -T) and β -Carotene (β -C), are aimed to protect, repair, and remove of the radical species generated. The LR^{\bullet}/α -T ratio was established as an appropriate indicator of oxidative condition in the hydrophobic cellular space (Malanga *et al.*, 2009). Considering the increasing awareness of the environmental risks generated by the use of GLY and its commercial formulations, the cellular pathways that could be affected in non-target organism by the exposure to the herbicide should be further studied. The hypothesis of the present work is that oxidative alterations could affect *C. vulgaris* redox balance in the hydrophobic cellular medium, consuming lipid soluble antioxidants, after the exposure to either pure GLY or the commercial formulate RUP. The effect of GLY and its formulate was measured on the biomass and rate of growth of the cultures, the oxidation rate of DCFH-DA, the generation rate of LR^{\bullet} , and the content of lipid soluble antioxidants (α -T and β -C).

Materials and Methods

Culture conditions and experimental design

Stock axenic cultures of *Chlorella vulgaris* CPCC90 (Origin: Canadian Phycological Culture Centre, strain identification: C90) were grown in sterilized Bold's Basal medium supplemented with 1 g l^{-1} glucose (Bold and Wynne, 1978). Cells were grown at $20 \pm 1^{\circ}\text{C}$ under light/dark cycles of 12:12 h. The irradiance at the surface of the culture was approximately 38 W m^{-2} of photosynthetically active radiation (PAR, Philips 50-W day-light-fluorescent light). Cell abundance was monitored by spectrophotometry ($\lambda = 600 \text{ nm}$) and cell counting using a hemocytometer. For cell counting, samples were taken and kept in plastic tubes with a solution of formaldehyde in water 10% (v/v).

Cells were grown in Bold's Basal medium with 1 g l^{-1} of glucose (control cells). For the treatments, the media was supplemented with either analytical grade GLY (Sigma-Aldrich) or a GLY-based herbicide (RoundUp UltraMax[®], RUP, donated by colleagues) to a final concentration of active ingredient (GLY) of $0.3 \mu\text{M}$ or $5 \mu\text{M}$ ($50 \mu\text{g l}^{-1}$ and $800 \mu\text{g l}^{-1}$, respectively). All determinations were done in exponential (Exp) (day 7) and/or stationary (St) (day 14) growth phases. In each phase the cells were harvested by centrifugation at $9,300 \text{ g}$ at 4°C . Cells were preserved at -80°C until the time of processing.

The diagram in Fig. 1 shows the experimental protocol designed for these studies.

Oxidation rate of DCFH-DA

Cellular cultures containing at least 2×10^6 cells ml^{-1} were used for assaying the oxidation rate of the dye. The harvested cells were resuspended using $400 \mu\text{l}$ of a Tris-HCl 40 mM pH 7.4 buffer. Two subsamples of this suspension containing $200 \mu\text{l}$ each were taken and transferred to new plastic tubes, then centrifuged at $9,280 \text{ g}$ at 4°C for 10 min. The supernatant was discarded and $100 \mu\text{l}$ of the same buffer were added to the pellet and resuspended. This new suspension was sonicated for 30 s using 10 s continuous pulses. An aliquot of $96 \mu\text{l}$ of buffer and $4 \mu\text{l}$ of DCFH-DA solution in methanol (1 mg ml^{-1}) was added to one of the subsamples. The sample blank was obtained by adding $100 \mu\text{l}$ of buffer to the other subsample. Reactant blank was prepared using the same proportion of buffer:DCFH-DA as the samples and their blanks. Both, the samples and the blanks, were incubated for 45 min at 30°C , followed by centrifugation at $9,280 \text{ g}$ for 5 min at 4°C . Supernatants were transferred to a 96-wells plaque. Fluorescence was determined at $\lambda_{\text{ex}} = 488 \text{ nm}$ and $\lambda_{\text{em}} = 525 \text{ nm}$ using a Varioscan Lux microplate reader.

Generation rate of LR^{\bullet}

The generation rate of LR^{\bullet} was determined by electronic paramagnetic resonance (EPR) using α -(4-pyridyl N-oxide)-N-tert-butylnitron (POBN) as a spin trap. Harvested cells were resuspended in $400 \mu\text{l}$ of a 0.5 M $\text{KH}_2\text{PO}_4/\text{K}_2\text{HPO}_4$ pH 7.0 buffer and sonicated for 15 s. An aliquot of $200 \mu\text{l}$ of the resulting homogenate was mixed with an equal volume of a 0.5 M solution of POBN prepared in the same buffer. Samples were placed in a glass micropipette and incubated

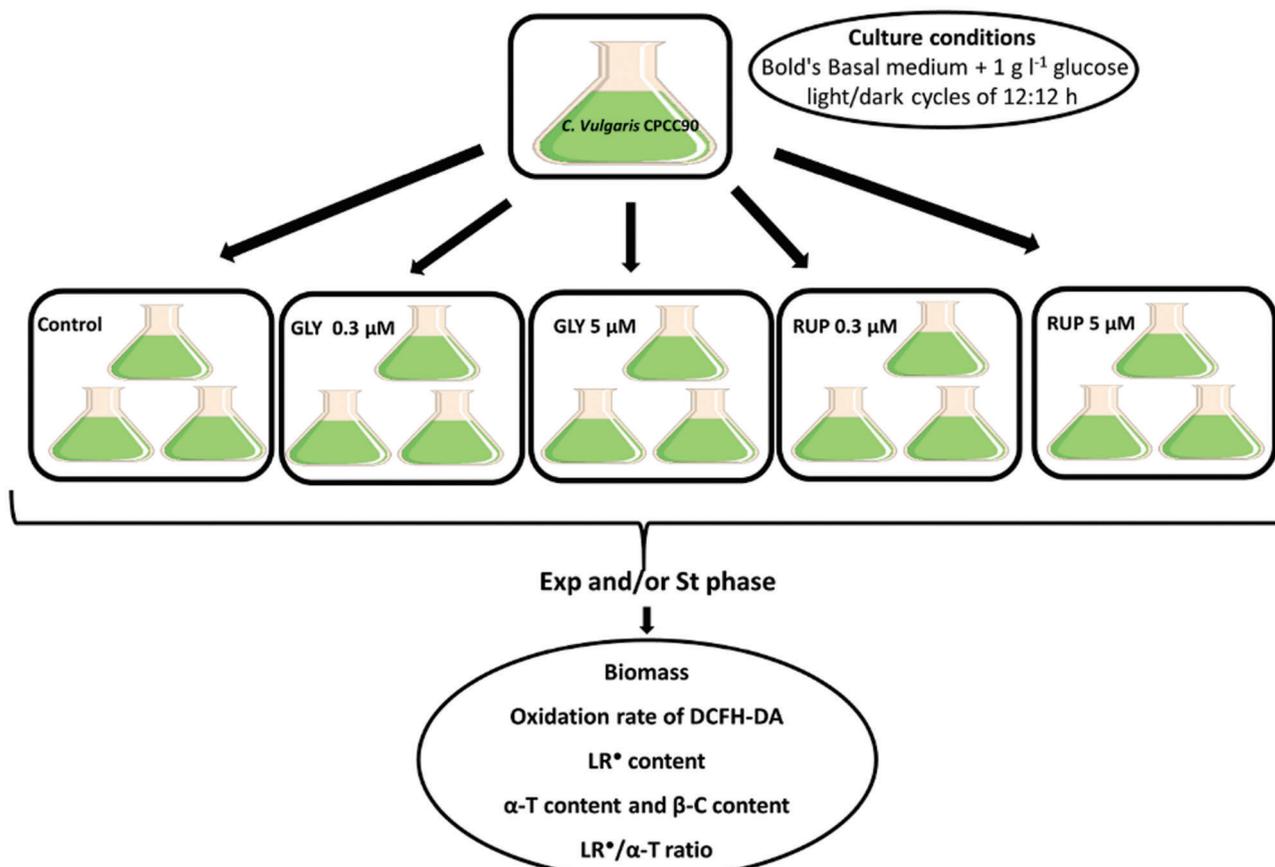


FIGURE 1. Diagram illustrating the experimental design used in this work.

at 30°C for 30 min before LR[•] determination. Spectra were measured using a Bruker (Karlsruhe, Germany) spectrometer ECS 106 with a cavity ER 4102ST, operating with the following instrument settings: 18°C room temperature, 50 kHz modulation frequency, 20 mW microwave power, 1 G modulation amplitude, 81.92 ms time constant and 2×10^4 receiver gain (Buettner and Jurkiewicz, 1993). Quantification of the spin adduct was performed using an aqueous solution of 4-hydroxy-2,2,6,6-tetramethylpiperidinyloxy (TEMPOL) introduced into the same sample cell and subjected to the same treatment as the samples. EPR spectra for both sample and TEMPOL solutions were recorded at exactly the same spectrometer settings and the first derivative EPR spectra were double integrated to obtain the area intensity, from which the content of the radical was calculated according to Kotake *et al.* (1996).

Content of lipid soluble antioxidants

The content of α-T and β-C was determined using high performance liquid chromatography (HPLC) with electrochemical detection in isocratic conditions according to Desai (1984) with modifications. Harvested cells were resuspended in 250 μl of a 1:1 methanol:ethanol solution and sonicated for 1 min for the extraction of the antioxidants. Afterwards, samples were kept for 30 min in the dark at 4°C. This procedure was repeated after the addition of 1 ml of hexane. After the last 30 min of incubation with hexane, samples were centrifuged at 2,320 g for 10 min at 4°C. An aliquot of 750 μl of the organic phase

was then transferred to a new plastic tube and evaporated to dryness using gaseous N₂ at room temperature. This extract was kept at -80°C until the time of injection. The HPLC procedure was performed using a Perkin-Elmer LC Binary Pump 250, ESA Coulochem II detector, ESA Model 5011 analytical cell, ESA Model 5020 Guard Cell, Perkin Elmer Model 1022 integrator and a manual injector with a 20 μl loop. The column employed was a Supelcosil LC-8 (150 mm × 4.6 mm; pore size 3 μm). Methanol and a 20 mM solution of lithium perchlorate in distilled water in a 97:3 proportion was used as the mobile phase. Flux was set to 1 ml min⁻¹ for a total run time of 13 to 15 min. Samples were resuspended using 150 μl of a 1:1 methanol:ethanol mixture and filtered using Nanosep MF 0.2 μm filters previous to injection. Quantification of antioxidants was made by injecting solutions of analytical standards of known concentration.

Statistical analysis

All statistical analyses were performed using the software GraphPad Prism 6.0. The results are expressed as mean ± ESM of at least three independent experiments, with three replicates in each experiment and were analyzed using 1-way or 2-way ANOVA with $p = 0.05$ or $p = 0.1$, unless otherwise specified.

Results

The growth cycle of *C. vulgaris* microalgae comprises three phases: lag phase (days 0–4), Exp phase (days 4–10) and St phase (days 11–14). To evaluate the effects of GLY and

RUP treatment on the biomass of the cultures, the cell count of the cultures was followed over 14 days, as shown in Fig. 2. A significant 42% decrease in the cellular biomass in St phase was observed in cultures supplemented with either 5 μM of GLY or RUP, as compared to control cultures (Table 1).

Growth rate (μ) was calculated according to Tajnaiová et al. (2020). The growth rate of cultures exposed to either 5 μM of GLY or RUP in Exp phase of development showed a non-significant decrease, as compared to control cultures (0.16 ± 0.04 , 0.4 ± 0.1 and 0.7 ± 0.2 10^7 cells ml^{-1} day $^{-1}$, respectively). Biomass was not affected after exposure of the cells to either 5 μM of GLY or RUP in Exp phase of development in agreement with previous reports (Ostera et al. (2020); however, a significant decrease in biomass, evaluated at day 14 in cultures treated with either GLY or RUP, was observed as compared to control cultures (Table 1).

Fig. 3 shows the DCFH-DA oxidation rate in control and GLY- and RUP-treated cultures in Exp and St growth phase. The treatment with 0.3 μM of GLY generated non-significant effects on the oxidation rate of DCFH-DA, neither in Exp nor in St phase of development as compared to control cultures. However, the treatment with GLY 5 μM lead to a significant decrease (62%) in the DCFH-DA oxidation rate as compared to control cultures either in Exp and St phase of cellular development, respectively. *C. vulgaris* cultures showed a significant decrease in the DCFH-DA oxidation rate after supplementation with either

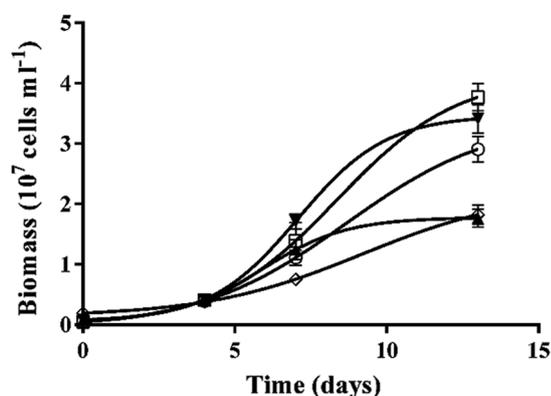


FIGURE 2. Kinetic of the biomass of the cultures of *C. vulgaris*. Biomass of Control cells (○); and cells grown in the presence of GLY 0.3 μM (□); GLY 5 μM (◇); RUP 0.3 μM (▼); RUP 5 μM (▲).

TABLE 1

Biomass in Exp and St phase of control, GLY- and RUP-treated cultures

Treatment	Biomass (10^7 cells ml^{-1})	
	Exp phase	St phase
Control	1.3 ± 0.2	3.1 ± 0.2
GLY 0.3 μM	1.4 ± 0.3	3.7 ± 0.2
GLY 5 μM	1.1 ± 0.2	$1.8 \pm 0.2^*$
RUP 0.3 μM	1.5 ± 0.2	3.4 ± 0.2
RUP 5 μM	1.2 ± 0.2	$1.8 \pm 0.1^*$

Note: *significantly different from control cultures (one-way ANOVA, $p < 0.05$).

0.3 μM or 5 μM of RUP in both phases of development. In cultures treated with 0.3 μM of RUP a significant 45% and 56% decrease in the oxidation rate was observed in Exp and St phase as compared to control cells, respectively. In the presence of 5 μM RUP, a significant decrease in the oxidation rate of the dye was observed in Exp and St phase compared to control cultures (52, 47%, respectively).

The LR $^{\bullet}$ generation rate for control and GLY- or RUP-treated cultures is shown in Fig. 4. The LR $^{\bullet}$ generation rate in Exp growth phase was not significantly affected by the supplementation of cultures with either 0.3 μM of GLY or RUP. However, a significant increase (23%) in the oxidation rate was observed only in cultures supplemented with 5 μM RUP, as compared to control in St phase.

Data in Table 2 shows the content of lipophilic antioxidants, α -T and β -C, in control, GLY- and RUP-treated cultures in both phases of development. The content of β -C was not significantly affected, as compared to control cultures, by any of the treatments in both growth phases of cellular development. In Exp phase, the content of α -T was significantly affected by the supplementation with 0.3 μM and 5 μM of RUP, showing an 87 and 84% decrease, as compared to control cultures, respectively. In St growth phase supplementation with either 0.3 or 5 μM GLY generated a significant 28 and 65% decrease in the α -T content, as compared to control cultures in the same growth phase, respectively. A significant 92% decrease was determined in the α -T content in cultures treated with either 0.3 or 5 μM of RUP, as compared to control cultures in this growth phase.

The LR $^{\bullet}$ / α -T content ratio, considered as an indicator of the balance between free radical damage and antioxidant protection in the hydrophobic medium (Malanga et al., 2009), was significantly different between samples obtained from control and RUP-exposed microalgae in both, Exp and St phase of development at either 0.3 or 5 μM formulate (Fig. 5). However, the ratio was not affected when *C. vulgaris* cells were exposed to 0.3 μM GLY in the Exp phase of development (Fig. 5 Insert), showing a more drastic effect of RUP, as compared to GLY tested under the same experimental conditions.

Discussion

GLY can affect non-target organisms in the aquatic environment, including microalgae and cyanobacteria, which share photosynthetic routes similar to those described for terrestrial plants (Hernández-García and Martínez-Jerónimo, 2020). They are not only affected in a similar fashion the shikimate pathway, but they also experience oxidative stress as well (Ostera et al., 2016). It was suggested than in addition to producing physiological and metabolic damage that impairs algal growth (Wu et al., 2016), GLY disrupts the integrity of the cell membrane (Amorós et al., 2007). A recent report (Hernández-García and Martínez-Jerónimo, 2020) evaluated the growth of an experimental community of chlorophycean microalgae (*Ankistrodesmus falcatus*, *C. vulgaris*, *Pseudokirchneriella subcapitata*, and *Scenedesmus incrassatulus*), by comparing their growth rate in the presence of the herbicide Faena $^{\circ}$ (1.564 mg GLY/l

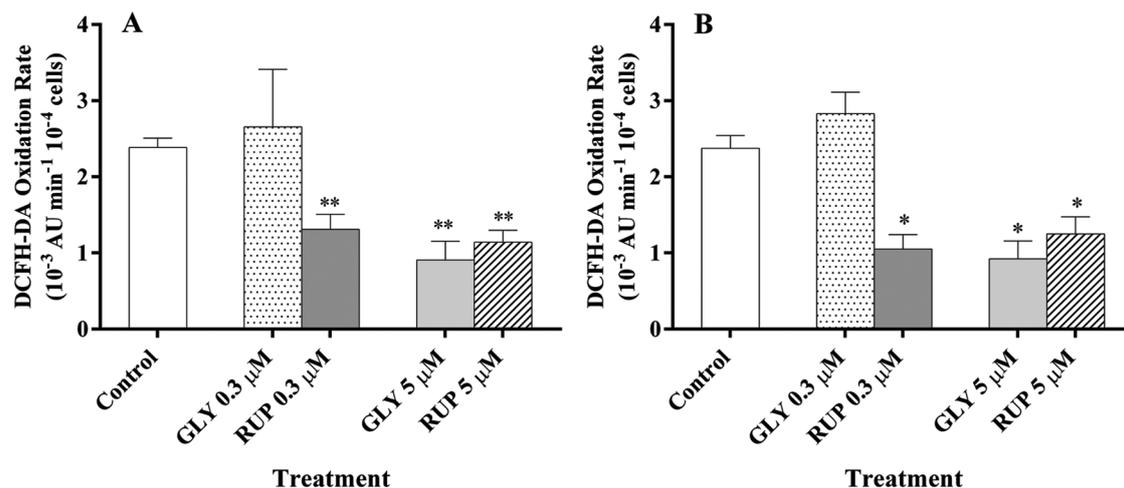


FIGURE 3. Oxidation rate of DCFH-DA of the cultures of *C. vulgaris* in the presence of either GLY or RUP in Exp (A) and St (B) growth phase of development. *significantly different from control cultures in the same growth phase (one-way ANOVA, $p < 0.05$). **significantly different from control cultures in the same growth phase (one-way ANOVA, $p < 0.1$).

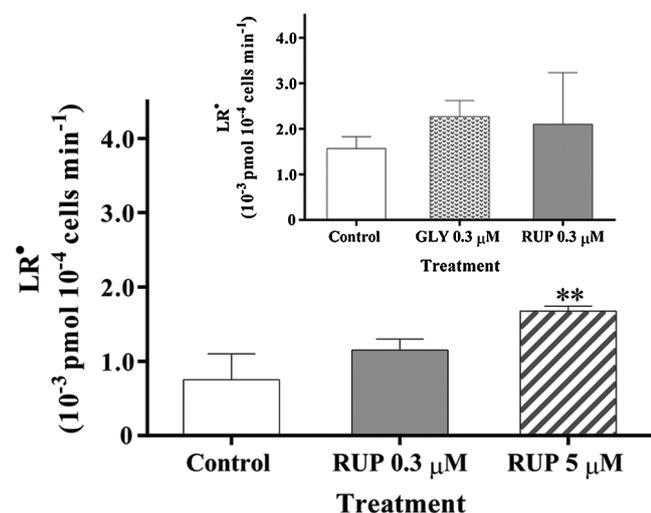


FIGURE 4. LR• generation rate in Control and GLY- RUP-treated cultures in Exp (inset) and St (main figure) growth phase. **significantly different from control cultures in the same growth phase (one-way ANOVA, $p < 0.1$).

during 14 days for *C. vulgaris*). In that study, negative, significant effects on the growth and physiology of green microalgae (revealed through the assessment of macromolecular and enzymatic biomarkers) were documented. In the same direction, the results reported here employing GLY (either pure or as RUP formulate), in a lower dose than the IC40 used previously, is in agreement with those reports since biomass of the cultures of *C. vulgaris* was significantly decreased in St phase of development. The observed effect in microalgae may result from the inhibition of the enzyme 5-enolpyruvylshikimate-3-phosphate synthase (EPSPS) in the shikimate pathway (Giesy *et al.*, 2000; Pérez *et al.*, 2011) and the possible oxidative stress triggered by the exposure to the herbicide, causing an imbalance in the enzymatic response (Ostera *et al.*, 2016).

The oxidation rate of DCFH-DA was successfully used as a general index of ROS generation in algae exposed to UV-B radiation (Malanga and Puntarulo, 1995), and in soybean roots *in vivo* exposed to Fe overload, and as an *in vitro* assay to study the ability of isolated microsomal membranes

TABLE 2

Content of α -T and β -C in control, GLY and RUP-treated cultures in Exp and St growth phase

	α -T content (pmol 10 ⁻⁴ cells)	β -C content (pmol 10 ⁻⁴ cells)
Exp phase		
• Control	0.4 ± 0.1	0.05 ± 0.02
• GLY 0.3 μM	0.3 ± 0.2	0.04 ± 0.03
• GLY 5 μM	0.19 ± 0.03	0.03 ± 0.02
• RUP 0.3 μM	0.05 ± 0.02**	0.006 ± 0.003
• RUP 5 μM	0.061 ± 0.004*	0.012 ± 0.004
St phase		
• Control	0.75 ± 0.07	0.04 ± 0.02
• GLY 0.3 μM	0.54 ± 0.05**	0.03 ± 0.01
• GLY 5 μM	0.26 ± 0.08*	0.03 ± 0.01
• RUP 0.3 μM	0.06 ± 0.01**	0.005 ± 0.002
• RUP 5 μM	0.06 ± 0.01*	0.005 ± 0.002

Note: *significantly different from control cultures in the same growth phase (one-way ANOVA, $p \leq 0.05$). **significantly different from control cultures in the same growth phase (one-way ANOVA, $p \leq 0.1$).

to generate superoxide anion (O_2^-) in the presence of NADPH (Caro and Puntarulo, 1996). Even more, Patetsini *et al.* (2013) examined, employing flow cytometry, the influence of environmentally relevant concentrations of two pesticides (chlorpyrifos and penoxsulam) on mussel physiological status. These works established this methodology as a reliable biomarker for the evaluation of pollution or other environmental stressors. The data reported here indicated that after the exposure to a low GLY concentration (0.3 μM) the oxidation rate of DCFH-DA was not affected neither in Exp nor in St phase of development of *C. vulgaris* cells. At

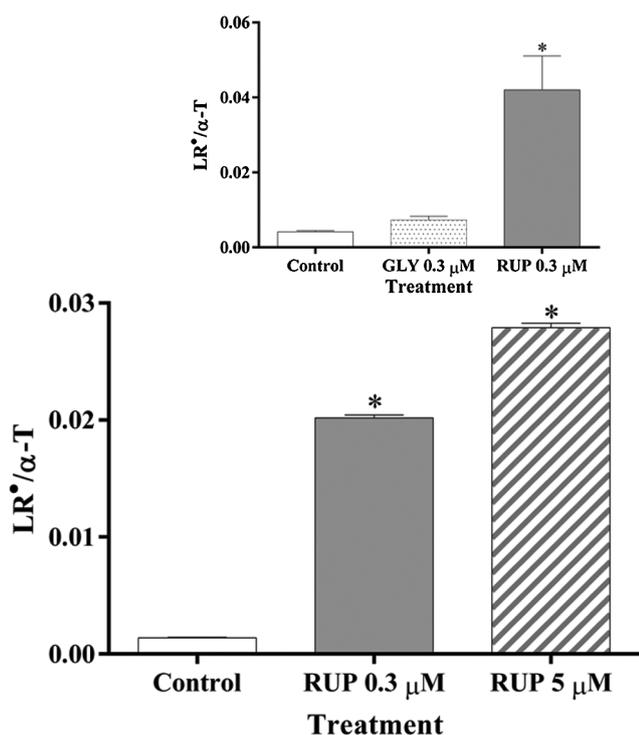


FIGURE 5. LR•/α-T ratio in Control and GLY/RUP treated cultures in Exp (inset) and St (main figure) growth phase. *significantly different from control cultures in the same growth phase (one-way ANOVA, $p < 0.05$).

the highest GLY concentration (5 μM) the oxidation rate of the dye was significantly decreased. This decrease was also detected after treatment with RUP at the studied concentrations. This unexpected response could be due to different possible effects of the herbicide:

(a) GLY that could be degraded by processes, such as metal oxides/UV, Fenton/UV, Fenton-like/UV, ozonation, chlorination (Yang *et al.*, 2018) was described to split into PO₄³⁻ and sarcosine by the chemisorption of H₂O and •OH in the existence of Fe oxides (Jaisi *et al.*, 2016). For the sarcosine path, it can react with •OH and O₂⁻ to form glycine and formaldehyde directly. Formaldehyde would be finally oxidized to formic acid and mineralized to CO₂ and water. Meanwhile, the decarboxylation of glycine might lead to the formation of methylamine due to the existence of •OH (Hidaka *et al.*, 1997). Through the oxidation of •OH and O₂⁻, methylamine could be further oxidized to ammonium (NH₄⁺), nitrite (NO₂⁻) and nitrate (NO₃⁻) (Manassero *et al.*, 2010). EPR studies verified the existence of •OH and O₂⁻ in both goethite/UV and magnetite/UV systems, and •OH dominated the photo-degradation of GLY (Yang *et al.*, 2018). In addition, as a highly reactive oxidizing agent, •OH and O₂⁻ could oxidize GLY to amino acid and carboxylic acid, and finally mineralized to inorganic ions. Overall, an effective photo-degradation of GLY in the magnetite/UV system was ascribed to a high capacity of generating ROS, crucially caused by a larger number of dissolved Fe ions. Fe oxides are the GLY scavengers in soils due to its high affinity with the phosphonate group of GLY (Borggaard and Gimsing, 2008). The sequestration of GLY on Fe oxides has been extensively

studied, but the knowledge on the subsequent GLY degradation on Fe oxides is limited. Due to the extensive existence of goethite and magnetite in nature, it should have a great impact on the fate of pesticides and other organic pollutants in the environment. As ROS can be formed *in situ* on a mineral surface, it can be reasonably inferred that a similar degradation reaction, described in the environment, will occur in the cells. Thus, the data reported here showing that the amount of reactive ROS available to react with the dye was decreased in the presence of GLY, as compared with the magnitude in its absence, could suggest that similar reactions to those reported in the cellular environment could occur in the cells, and that the presence of Fe is key factor in this scenario. However, until now this mechanism was not studied in biological organisms and further research in this field is strongly required.

(b) Lipid soluble antioxidants are the best candidates to be involved in the protection mechanisms against GLY-dependent damage since oxidative balance in the hydrophilic medium of *C. vulgaris* cells was affected by this treatment (Ostera *et al.*, 2020). The data reported here clearly suggested that oxidative redox balance was altered at the lipophilic cellular media, such as indicated by the increase in the LR• generation rate and in the LR•/α-T ratio, after exposure of the microalgae to RUP.

The content of β-C, that is very low in the cell as compared to other antioxidants, did not change by GLY administration. Tocopherol biosynthesis involves two metabolic pathways: the first is via homogentistic acid through the cytosolic shikimate pathway whereas the second is the plastid methylerythritol phosphate pathway for the tocopherol tail (phytyl diphosphate) synthesis. Finally, the assembly of the aromatic ring and phytyl diphosphate tail gives rise to tocopherols (Herrmann and Weaver, 1999). The significant decrease in the content of α-T observed in *C. vulgaris* cells supplemented with either GLY or RUP, strongly suggested that the consumption of this antioxidant is an important defense barrier to the deleterious effect of the herbicide. On the other hand, α-T works synergistically with other antioxidant species such as carotenoids, glutathione (GSH), and ascorbate (AH-) and preserve plastid redox homeostasis in stressed plants (Munné-Bosch, 2005). The formation of these conjugated low molecular weight antioxidants protects further plant metabolism against stress injuries (Szarka *et al.*, 2012). There are two possible means of enhancing conjugated antioxidant formation, increasing either the biosynthesis or the redox cycling of these compounds, and both can result in an increased biological activity of the compounds. In the cell, the interdependence of these antioxidants could play a basic role in electron flow (Gill and Tuteja, 2010).

Synthesis of various aromatic amino acids and precursors of numerous secondary metabolites, like vitamins and pigments takes place through the shikimate pathway (Herrmann and Weaver, 1999). In this pathway, one of the glycolytic cycle compounds, phosphoenol pyruvic acid, and another transitional compound, erythrose-4-phosphate are transformed into chorismate. The reversible synthesis of 5-enolpyruvylshikimate 3-phosphate and phosphoenolpyruvate is mediated by the enzyme 5-enolpyruvylshikimate 3-phosphate synthase (Velisek and Cejpek, 2006) and inhibited by GLY (Lushchak and Semchuk, 2012). Thus, it can be postulated that

the more drastic decrease in α -T content observed in the St phase. as compared to Exp phase, could be due to this inhibitory effect of GLY in the synthesis of the antioxidant required as the growth of the culture proceed.

(c) A contribution of the already mentioned effects (a) and (b) to the decrease of the oxidation rate of the dye in cells exposed to either GLY or RUP.

The abundant and controversial studies on GLY effect on photosynthetic organisms is a clear indication of the complexity of the scenario faced when oxidative metabolism, in both target and non-target plants, is characterized. Further analysis is still required to solve this point. The data presented here are experimental evidence of the contribution of GLY-dependent effects on redox balance in the lipophilic cellular media, that is reflected in the consumption of endogenous antioxidant α -T leading to a potential decrease of the cells ability to afford further oxidative challenges. Moreover, this study shows that the formulate RUP is able to damage the cellular membranes causing deleterious effects on the cells. This is a critical issue for the agrochemical use of the herbicide, since the addition of surfactants is very much used to enhance the efficiency of the product. Thus, characterization of the RUP effects will be encouraged to provide a tool to find new alternatives for its use and, to assess the real contribution of GLY to the reported effects after exposure in the natural environment.

Availability of Data and Materials: All data generated or analyzed during this study are included in this published article (and its supplementary information files).

Author Contribution: GM and SP participated in the designing of the original idea and in the manuscript writing. JMO contributed to experimental work, the bibliographic search and in the manuscript writing. All authors have approved the final article.

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Conflicts of Interest: GM and SP participated in the designing of the original idea and in the manuscript writing. JMO contributed to experimental work, the bibliographic search and in the manuscript writing. All authors have approved the final article.

References

- Ahsan N, Lee DG, Lee KW, Alam I, Lee SH et al. (2008). Glyphosate-induced oxidative stress in rice leaves revealed by proteomic approach. *Plant Physiology and Biochemistry* **46**: 1062–1070.
- Altamar Ríos J (2007). Factores ambientales de riesgo, radicales libres y enfermedad. *Revisión. Revista Electrónica de PortalesMedicos.com. Medicina Preventiva y Salud Pública*, **2**: 154–155.
- Amorós I, Alonso JL, Romaguera S, Carrasco JM (2007). Assessment of toxicity of a glyphosate-based formulation using bacterial systems in lake water. *Chemosphere* **67**: 2221–2228.
- Bold HC, Wynne MJ (1978). Cultivation of algae in the laboratory. In: McElroy WD, Swanson CP (eds.), *Introduction to the Algae Structure and Reproduction*. Englewood Cliffs, NJ: Prentice Hall, 571.
- Borggaard OK, Gimsing AL (2008). Fate of glyphosate in soil and the possibility of leaching to ground and surface waters: A review. *Pest Management Science* **64**: 441–456.
- Buettner GR, Jurkiewicz BA (1993). Ascorbate free radical as a marker of oxidative stress: An EPR study. *Free Radical Biology and Medicine* **14**: 49–55.
- Cakmak I, Yazici A, Tutus Y, Ozturk L (2009). Glyphosate reduced seed and leaf concentrations of calcium, manganese, magnesium, and iron in non-glyphosate resistant soybean. *European Journal of Agronomy* **31**: 114–119.
- Caro A, Puntarulo S (1996). Effect of *in vivo* iron supplementation on oxygen radical production by soybean roots. *Biochimica et Biophysica Acta* **1291**: 245–251.
- Chen FX, Zhou CR, Li GP (2012). Study on thermal decomposition and the non-isothermal decomposition kinetics of glyphosate. *Journal of Thermal Analysis and Calorimetry* **109**: 1457–1462.
- Choi CJ, Berges JA, Young EB (2012). Rapid effects of diverse toxic water pollutants on chlorophyll a fluorescence: variable responses among freshwater microalgae. *Water Research* **46**: 2615–2626.
- Desai ID (1984). Vitamin E analysis methods for animal tissues. *Methods in Enzymology* **105**: 138–147.
- Duke SO, Lydon J, Koskinen WC, Moorman TB, Chaney RL, Hammerschmidt R (2012). Glyphosate effects on plant mineral nutrition, crop rhizosphere microbiota, and plant disease in glyphosate-resistant crops. *Journal of Agricultural and Food Chemistry* **60**: 10375–10397.
- Eker S, Ozturk L, Yazici A, Erenoglu B, Romheld V et al. (2006). Foliar-applied glyphosate substantially reduced uptake and transport of iron and manganese in sunflower (*Helianthus annuus* L.) plants. *Journal of Agricultural and Food Chemistry* **54**: 10019–10025.
- Giesy JP, Dobson S, Solomon KR (2000). Ecotoxicological risk assessment for Roundup® herbicide. *Reviews of Environmental Contamination and Toxicology* **167**: 35–120.
- Gill SS, Tuteja N (2010). Reactive oxygen species and antioxidant machinery in abiotic stress tolerance in crop plants. *Plant Physiology and Biochemistry* **48**: 909–930.
- Gomes MP, Juneau P (2016). Oxidative stress in duckweed (*Lemna minor* L.) induced by glyphosate: Is the mitochondrial electron transport chain a target of this herbicide? *Environmental Pollution* **218**: 402–409.
- Gomes MP, Le Manac’h SG, Maccario S, Labrecque M, Lucotte M et al. (2016a). Differential effects of glyphosate and aminomethylphosphonic acid (AMPA) on photosynthesis and chlorophyll metabolism in willow plants. *Pesticide Biochemistry and Physiology* **130**: 65–70.
- Gomes MP, Le Manac’h SG, Moingt M, Smedbol E, Paquet S et al. (2016b). Impact of phosphate on glyphosate uptake and toxicity in willow. *Journal of Hazard Materials* **304**: 269–279.
- Hernández-García CI, Martínez-Jerónimo F (2020). Multistressor negative effects on an experimental phytoplankton community. The case of glyphosate and one toxigenic cyanobacterium on Chlorophycean microalgae. *Science of the Total Environment* **717**: 137–186.
- Herrmann KH, Weaver LM (1999). The Shikimate pathway. *Annual Review of Plant Physiology and Plant Molecular Biology* **50**: 473–503.

- Hidaka H, Horikoshi S, Ajisaka J, Zhao J, Serpone N (1997). Fate of amino acids upon exposure to aqueous titania irradiated with UV-A and UV-B radiation Photocatalyzed formation of NH_3 , NO_3^- , and CO_2 . *Journal of Photochemistry and Photobiology A: Chemistry* **108**: 197–205.
- Kotake Y, Tanigawa T, Tanigawa M, Ueno I, Allen DR, Lai C (1996). Continuous monitoring of cellular nitric oxide generation by spin trapping with an ironthiocarbamate complex. *Biochimistry and Biophysical Acta* **1289**: 362–368.
- Jaisi DP, Li H, Wallace AF, Paudel P, Sun MJ, Balakrishna A, Lerch RN (2016). Mechanisms of bond cleavage during manganese oxide and UV degradation of glyphosate: results from phosphate oxygen isotopes and molecular simulations. *Journal of Agricultural and Food Chemistry* **64**: 8474–8482.
- Lipok J, Studnik H, Gruyaert S (2010). The toxicity of Roundup® 360 SL formulation and its main constituents: Glyphosate and isopropylamine towards nontarget water photoautotrophs. *Ecotoxicology and Environmental Safety* **73**: 1681–1688.
- Lushchak VI, Semchuk NM (2012). Tocopherol biosynthesis: Chemistry, regulation and effects of environmental factors. *Acta Physiologiae Plantarum* **34**: 1607–1628.
- Malanga G, Puntarulo S (1995). Oxidative stress and antioxidant content in *Chlorella vulgaris* after exposure to ultraviolet-B radiation. *Physiologia Plantarum* **94**: 672–679.
- Malanga G, Perez A, Calvo J, Puntarulo S (2009). The effect of seasonality on oxidative metabolism in the sea urchin *Loxechinus albus*. *Marine Biology* **156**: 763–770.
- Manassero A, Passalia C, Negro AC, Cassano AE, Zalazar CS (2010). Glyphosate degradation in water employing the $\text{H}_2\text{O}_2/\text{UVC}$ process. *Water Research* **44**: 3875–3882.
- Munné-Bosch S (2005). The role of α -tocopherol in plant stress tolerance. *Journal of Plant Physiology* **162**: 743–748.
- Ostera JM, Malanga G, Puntarulo S (2016). Actualización sobre aspectos oxidativos del efecto del glifosato en sistemas biológicos/An update on the effects of glyphosate on the oxidative state in biological systems. *Biotecnología* **18**: 3–10.
- Ostera JM, Malanga G, Puntarulo S (2020). Assessment of oxidative balance in hydrophilic cellular environment in *Chlorella vulgaris* exposed to glyphosate. *Chemosphere* **248**: 125955.
- Patetsini E, Dimitriadis VK, Kaloyianni M (2013). Biomarkers in marine mussels, *Mytilus galloprovincialis*, exposed to environmentally relevant levels of the pesticides, chlorpyrifos and penoxsulam. *Aquatic Toxicology* **126**: 338–345.
- Pérez GL, Vera MS, Miranda LA (2011). Effects of herbicide glyphosate and glyphosate-based formulations on aquatic ecosystems. *Herbicides and Environment* **16**: 343–368.
- Qian H, Chen W, Sun L, Jin Y, Liu W, Fu Z (2009). Inhibitory effects of paraquat on photosynthesis and the response to oxidative stress in *Chlorella vulgaris*. *Ecotoxicology* **18**: 537–543.
- Reno U, Regaldo L, Vidal E, Mariani M, Zalazar C, Gagneten AM (2016). Water polluted with glyphosate formulations: effectiveness of a decontamination process using *Chlorella vulgaris* growing as bioindicator. *Journal of Applied Phycology* **28**: 2279–2286.
- Romero DM, de Molina MCR, Juárez AB (2011). Oxidative stress induced by a commercial glyphosate formulation in a tolerant strain of *Chlorella kessleri*. *Ecotoxicology and Environmental Safety* **74**: 741–747.
- Szarka A, Tomasskovic B, Bánhegyi G (2012). The ascorbate-glutathione- α -tocopherol triad in abiotic stress response. *International Journal of Molecular Sciences* **13**: 4458–4483.
- Tajnaiová L, Vurm R, Kholomyeva M, Kobera M (2020). Determination of the Ecotoxicity of Herbicides Roundup® Classic Pro and Garlon New in aquatic and terrestrial environments. *Plants* **9**: 1203–1224.
- Velisek J, Cejpek K (2006). Biosynthesis of food constituents: Amino acids: 2. The alanine-valine-leucine, serine-cysteine-glycine, and aromatic and heterocyclic amino acids groups. *Czech Journal of Food Sciences* **24**: 45–58.
- Wu L, Qiu Z, Zhou Y, Du Y, Liu C, Ye J, Hu X (2016). Physiological effects of the herbicide glyphosate on the cyanobacterium *Microcystis aeruginosa*. *Aquatic Toxicology* **178**: 72–79.
- Yang Y, Deng Q, Yan W, Jing C, Zhang Y (2018). Comparative study of glyphosate removal on goethite and magnetite: Adsorption and photo-degradation. *Chemical Engineering Journal* **352**: 581–589.